

The Importance of Geochemistry in the Management of Lake Environments: Analysis of Parameters Composing the Water Quality Index (Wqi)*

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Abstract

The present work aims to elaborate a review of the geochemical aspects used to compose the Raw Water Quality Index (WQI), highlighting lake environments and their importance for management practices. This paper is a bibliographic review based on specialized scientific and legislative literature conducted between April 2016 and March 2019. The scientific articles were selected by searching the Scielo database and the Periódicos Capes and the database search was performed using the terminologies registered in the Descriptors (Thesaurus) or Keywords (Scopus and Web of Science) for geosciences and environmental sciences. The WQI is the main quality index used in the country and is calculated based on nine parameters (water temperature, pH, dissolved oxygen, total residue, biochemical oxygen demand, thermotolerant coliforms, total nitrogen, total phosphorus and turbidity). The values determined through the calculation of the WQI are classified in ranges that define the water quality, which varies between different Brazilian states. Among its mishaps, it is noteworthy that the WQI does not analyze other factors important for public supply, such as toxic substances (eg: heavy metals, pesticides, organic compounds), pathogenic protozoa and substances that interfere with the organoleptic properties of water, which are under other management tools. The systemic model of water resources management adopted for Brazil is still in the process of improvement, because it still works through a decentralized system where the public authorities, at different levels, have an active voice, thus requiring the commitment to make a long-term monitoring based on a reliable geochemical analysis, so that an integrated management can be carried out, allowing the synthesis and characterization of the current state and management criteria of lake environments, with the main objective of facing the conservation challenges allied to sustainable use of its resources.

Keywords

WQI, Lakes, managemet, parameters

Subject Areas: Environmental Sciences,Geochemistry,Hydrology

1 1. Introduction

2 Lakes and lagoons are ecosystems belonging to watersheds of recognized importance in different geological and
3 ecological processes, such as full and ebb attenuation, reduction of surface erosion, water quality conditioning and
4 maintenance of canals by the protection of margins and siltation reduction [1]. There is broad scientific consensus
5 (ABC/SBPC) that these are ecosystems that, for their stability and functionality, need to be conserved or restored if
6 historically degraded.

7 Degradation of lakes and lagoons in the world is an important topic for Global Aquatic Resources, which arouses
8 and requires attention, since lakes, in addition to their importance as a component of the ecosystem regulation, contains
9 more than 90% of the fresh water available for human consumption in terrestrial surface [2].

10 Among the main values of the resources offered by the lakes and reservoirs, the most relevant ones can be divided
11 into three main groups: provision services (drinking water supply, irrigation, navigation, fishing and tourism),
12 regulation services (flood and drought, management, drainage, climate remediation) and cultural services (religious and
13 historical values) [3].

14 With the continually increasing of global population, providing adequate drinking water supplies is becoming a
15 critical challenge. Therefore, there is a crucial importance regarding the attention of different administrative spheres to
16 the protection of lake and reservoir ecosystems, the former being more vulnerable to different environmental impacts.

17 Currently, Brazil has a solid legal framework for water management, especially the decentralized management that
18 allows the participation of the Public Authority at different levels of action, of users and society in general [4].

19 At the apex of the institutional framework is the National Water Resources Council and, in the states, the State
20 Water Resources Councils, which discipline procedures and arbitrate conflict situations. At the same time, there are
21 about 150 state and interstate basin committees installed in different regions of the country, which deliberate on water
22 resources plans and required action priorities [5].

23 This set, which configures the National Water Resources Management System - SINGREH, along with the
24 management instruments defined by law¹, are in full force throughout the country. Thus, the differences regarding the
25 institutionality, the management criteria, and the technical capacity installed in each state are clear, which, to some
26 extent, still require improvement [6].

27 The wide necessity and scope of geoscientific studies, and especially the geochemical ones, is justified not only for
28 the environmental scientific knowledge, but mainly in the determination of the impacts that the different types of uses,
29 enterprises, projects and actions in the lakes and lagoons can cause when effectively immersed in physical reality [7].

30 The parameters commonly used to study water quality in lakes and reservoirs are: (physical) sunlight, temperature,
31 turbidity, particulate matter in suspension, conductivity and color; (chemical) dissolved oxygen, salinity, pH, alkalinity,
32 phosphates, nitrates and ammonia, organic carbon, among others; (biotic) phytoplankton, zooplankton, fish, crustaceans,
33 aquatic invertebrates; (hydrological) surface area, discharge or flow, holding time, fluctuation, depth, continuity;
34 (geomorphological) vegetation cover of the basin, urban agglomerations, riparian and riparian forest, landscape
35 evolution [8].

36 According to the National Water Agency, the main water quality index used in the country is the WQI, created in
37 the United States by the *National Sanitation Foundation* in 1970 and used by CETESB (Environmental Company of the
38 State of São Paulo) since 1975. The WQI is a simple mathematical model that determines water quality based on nine
39 parameters (thermotolerant coliforms, pH, biochemical oxygen demand, nitrate, total phosphate, water temperature,
40 turbidity, total solids and dissolved oxygen), giving a corresponding weight for each according to their importance.

41 The WQI therefore bases its qualification on water quality based on parameters that bring as aggregate information
42 water contamination due to organic and fecal matter, solids and nutrients, but has limitations when considering that this
43 index does not take into account other parameters indispensable for the qualification of a body of water as suitable for
44 public supply, such as the incidence of toxic substances (e.g. heavy metals, pesticides, organic compounds), occurrence
45 of pathogenic protozoa and other substances that may interfere with organoleptic properties of water.

46 Nutrients are essential elements for the development of aquatic biota, especially in confined environments that can,
47 in excess, cause the eutrophication of lakes and reservoirs, that is, their exaggerated enrichment, with considerable
48 damage to the environment. The main nutrients are Phosphorus, Nitrogen, Carbon and Silica [8].

¹They are instruments of the National Water Resources Policy, according to Law no. 9.433 / 97: a) Water Resource Plans; b) the classification of water bodies into classes, according to the predominant uses of water; c) the granting of rights to use water resources; d) charging for the use of water resources; and e) the Water Resources Information System.

49 Therefore, this paper aims to elaborate a synthesis and characterization of the current state of management criteria
50 and review of geochemical aspects and environmental indicators evaluated in lake environments used to compose the
51 raw Water Quality Index (WQI).

52 2. Material and Methods

53 This paper is a bibliographic review based on specialized scientific and legislative literature conducted between
54 April 2016 and March 2019. The scientific articles were selected by searching the Scielo database and the Periódicos
55 Capes portal using the publication dates (considering the lower limit of 2010) and the impact factor (ratio between the
56 number of times articles of a journal are cited and the total number of articles published each year) as the criteria of
57 their choice in each journal consulted.

58 The database search was performed using the terminologies registered in the Descriptors (Thesaurus) or Keywords
59 (Scopus and Web of Science) for geosciences and environmental sciences.

60 The methodology of this research can be divided into two main stages: (i) a stage regarding the thorough search
61 for scientific texts and official legal clauses related to the subject; (ii) another in which the texts found were analyzed
62 regarding their content.

63 To conclude the analyses, discussion meetings were held with the Interface Geochemistry research group, from the
64 Federal University of Bahia.

65 3. Results

66 3.1 Management Criteria

67 The maximum levels of impurities allowed in water are set according to their uses. These constitute the quality
68 standards that are set in order to ensure that the water to be used does not contain impurities that may harm it or, in the
69 case of effluent discharges, that they will not cause damage to the environment.

70 It is important to highlight the difference between criteria and standards of water quality. The definition of water
71 quality criteria deals with maximum tolerable values that guarantee the intended uses of water considering generic
72 exposure conditions. The criterion becomes a quality standard when it is cited in legislation, its establishment requires
73 appropriate toxicological studies [3].

74 Water quality criteria are established individually for each type of use, namely: human consumption, recreation,
75 animal desedentation, irrigation, aquatic life protection and aquaculture.

76 Water quality standards express desirable physical, chemical and biological characteristics in waters due to their
77 predominant uses. In order to protect water bodies, legal restrictions have been instituted to maintain physical, chemical
78 and biological characteristics within certain limits [9].

79 Water quality standards and criteria in Brazilian legislation are produced with a view to meeting water uses, such
80 as public supply, bathing, irrigation, fishing, among others. Such uses are protected by the standards set by the
81 following laws: Ministry of Health Ordinance No. 2914/11, CONAMA Resolution No. 274/2000, CONAMA
82 Resolution No. 357/2005 [3].

83 The occurrence and management of problems in a lake environment are mainly influenced by three specific
84 characteristics of this type of ecosystem: (i) its integrative nature, (ii) long retention time and (iii) a complex response
85 dynamics [10].

86 The first feature means that problems in these environments are rarely located within lakes. Its origin comes from
87 the environment in which it is located and is eventually introduced as part of ecosystem problems, for example flooding
88 affecting the lake shore; pollution that spreads can affect much of the environment, and biological problems, such as
89 non-native species but eventually introduced, also pose threats. The relatively long retention time means that problems
90 take a long time to become apparent. And the complexity of lake dynamics influences the way problems become
91 evident.

92 Water quality is the result of processes that occur on the surface of the watershed and is strongly dependent on the
93 natural characteristics of the watershed, such as soil type, vegetation and climate, as well as on anthropic activities [3].

94 Among the main problems behind the preservation of today's lake basins of biological origin are the overgrowth of
95 aquatic plants, which alter native fauna habitats, interfere with water transport, harbor harmful species, prevent water

96 flows and increases lake surface evapotranspiration [11] and the introduction of wild species that can compete for food
97 resources or become predators of native species [21].

98 Regarding aspects of chemical origin, there are changes in the salinity of lakes that disrupt some ecosystems,
99 consequently harming the communities dependent on them. Nutrient accumulation is also a factor that influences the
100 various communities promoting the eutrophication of aquatic plants. This problem is particularly common in Asian
101 lakes. Soil erosion and fertilizer use are also responsible for excess diffuse inputs from nutrient sources, which can
102 cause so-called agro-chemical pollution, which affects aquatic food chains and makes fish unfit for human consumption
103 [11].

104 Human activities such as deforestation, unproductive land use and poor management of riparian forest in lake
105 basins result in excess sediment inputs and reduction of wetlands, not to mention dumping untreated or poorly treated
106 waste, contaminating lakes and reducing concentrations of oxygen and nutrients, which increases the eutrophication of
107 the lake. In addition to these factors, there are predictions that climate change and global warming cause changes in
108 precipitation and runoff, as well as changes in the thermodynamic and ecological balance of lakes [10].

109 Watershed and water resources problems are very similar around the world. What differs is the perception about
110 the seriousness of the problems and the ability of governments to deal with them. Eutrophication, for example, is seen
111 as a negative aspect in developed countries. However, the increase in aquatic production associated with eutrophication
112 is considered a positive quality in Southeast and South Asia and in other regions where aquaculture is practiced and fish
113 enclosures and farms exist [10].

114 3.2 Geochemical Parameters

115 Dissolved oxygen

116 Among the gases dissolved in water, oxygen (O₂) is one of the most important in the dynamics and
117 characterization of aquatic ecosystems. The main sources of oxygen for water are the atmosphere and photosynthesis.
118 On the other hand, the losses are the consumption by decomposition of organic matter (oxidation), losses to the
119 atmosphere, respiration of aquatic organisms and oxidation of metallic ions such as iron and manganese [12].

120 Dissolved Oxygen is indispensable to aerobic organisms; waters with low dissolved oxygen levels indicate that
121 they have received organic matter; decomposition of organic matter by aerobic bacteria is usually accompanied by
122 consumption and reduction of dissolved oxygen in water; depending on the self-purification capacity of the lake, the
123 dissolved oxygen content can reach very low values or zero, extinguishing aerobic aquatic organisms. Dissolved oxygen
124 is the main parameter for characterizing the effects of water pollution [3; 13].

125 The pattern of oxygen distribution in aquatic ecosystems is, as a rule, inverse to carbon dioxide. When intense
126 carbon dioxide consumption occurs in the euphotic zone due to photosynthesis, while considerable oxygen production
127 occurs. On the other hand, in the aphotic zone, due to microbial activity (decomposition of organic matter), there is a
128 high carbon dioxide production and corresponding oxygen consumption [14; 15].

129 The solubility of oxygen in water, as with all gases, depends on two main factors: temperature and pressure. Thus,
130 as the temperature rises and the pressure decreases, there is a reduction and solubility of oxygen in water [16; 17; 13].
131 Some indirect factors may also be cited as controlling for dissolved oxygen concentration, such as the length of the
132 thermal stratification period, the concentration of organic matter (dissolved and particulate) of the water, and the
133 phytoplankton photosynthesis rate. Based on these properties it is concluded that tropical aquatic organisms have in
134 principle less oxygen available than those of temperate lakes [12].

135 3.3 Thermotolerant Coliforms

136 Thermotolerant coliforms is one of the most widely used microbial indicators of water quality [18]. This parameter
137 is based on the development of *Enterobacteriaceae* family bacilli, of fecal or environmental origin, within the
138 temperature range $44.5 \pm 0.2^\circ\text{C}$. Tolerance within this temperature range determines the differentiation between fecal
139 (thermotolerant) and non-fecal (thermosensitive) subgroups [19; 20].

140 The *Enterobacteriaceae* family includes oxidase-negative, non-spore-forming bacteria of approximately 0.4-0.7
141 mm in size, characterized by the activity of the enzyme β -galactosidase (fermenting lactose at temperatures of 44° -
142 45°C , with production of acid, gas and aldehyde) [22]. Four groups of these bacteria, including *Escherichia coli*,
143 *Citrobacter*, *Enterobacter* and *Klebsiella*, are widely used as indicators of thermotolerant coliforms, distributed in
144 different proportions in different environments, such as water, sewage, soil, vegetables and food. Of these, only *E. coli*

145 is guaranteed to be present in humans and homeothermic animal feces with percentages of around 96 to 99%. The other
146 kinds participate with percentages ranging from 3 to 8% in animal feces to 3 to 4% in human feces [19; 20; 23].

147 The determination of the presence of thermotolerant coliforms is of great importance as an indicator of fecal
148 contamination, since even in waters of secondary and recreational contact, thermotolerant coliforms are considered to
149 cause waterborne pathologies (**otitis externa**, gastrointestinal disorders characterized by diarrhea, **urinary tract**
150 **infections**, pneumonia and meningitis), or indicative of the presence of other pathogens [24; 25].

151 3.4 Potential of Hydrogen (pH)

152 In pure water there is an albeit small amount of H and OH⁻ ions in equilibrium with the water molecules. The
153 basicity or acidity of a solution is often expressed in terms of ion concentration, H⁺ which is determined pH, which is
154 defined as the negative logarithm of the hydrogen ion molar concentration [12].

155 In a natural environment, many factors can influence pH values, which makes it, despite its importance, one of the
156 most difficult environmental variables to interpret. In most surface waters the pH is influenced by the concentration of
157 H ions from carbonic acid dissociation, which generates low pH values and the reactions of carbonate and bicarbonate
158 ions with the water molecule that raise the pH values to the alkaline range [26].

159 Most continental water bodies have a pH ranging from 6 to 8; however, one may encounter more acidic or more
160 alkaline environments. In both cases, these ecosystems have also characteristic plant and animal communities [9].

161 Low pH lakes are often found in volcanic regions. The low pH of these ecosystems (<2) is mainly due to the
162 presence of sulfuric acid and other mineral salts whose cations hydrolyze [12].

163 Aquatic ecosystems with high pH values are generally found in regions with negative water balance (where
164 precipitation is less than evaporation); in regions where continental aquatic ecosystems are, to varying degrees,
165 influenced by the sea (receive large contributions from carbonates and bicarbonates), and in karst regions (calcium rich
166 regions) [26].

167 The importance of pH as a limiting factor for colonization of aquatic ecosystems by different organisms was first
168 recognized by Thienemann (1918) [27]. This fact served as a basis for formulating a basic principle in ecology: “The
169 more the living conditions of a biotope deviate from the optimal conditions for most organisms, the poorer will be the
170 communities species, the more uniform and more typical they will be, and the greater the number of organisms in each
171 species”.

172 The pH also influences the water treatment steps (coagulation and level of scalability/corrosivity) and the control
173 of the operation of sewage treatment plants (anaerobic digestion).

174 3.5 Nitrogen

175 Nitrogen is one of the most important elements in the metabolism of aquatic ecosystems. This importance is
176 mainly due to its participation in protein formation, one of the basic components of biomass. When present in low
177 concentrations, it can act as a limiting factor in the primary production of aquatic ecosystems [28].

178 The main natural sources of nitrogen can be rain, organic and inorganic material and the fixation of molecular
179 nitrogen within the lake itself. The increase of nitrogen in water is associated with the presence of domestic, industrial
180 sewage and fertilizers. The predominant form of nitrogen may provide the pollution stage: recent pollution is associated
181 with organic nitrogen or ammonia, while more remote pollution is associated with nitrogen in the form of nitrate [3].

182 Nitrogen is present in aquatic environments in various forms, for example: nitrate (NO₃), nitrite (NO₂), ammonia
183 (NH₃), ammonium ion (NH₄⁺), nitrous oxide (N₂O), molecular nitrogen (N₂), dissolved organic nitrogen (peptides,
184 purines, amines, amino acids, etc.), particulate organic nitrogen (bacteria, phytoplankton, zooplankton and detritus), etc.
185 [12].

186 Among the different forms, nitrate, along with ammonium ion, play a great role in aquatic ecosystems, as they
187 represent the main sources of nitrogen for primary producers [29].

188 High concentrations of the ammonium ion may have major ecological implications, such as: strongly influencing
189 the dissolved oxygen dynamics of the environment and influencing the fish community, because at basic pH the
190 ammonium ion becomes ammonia (free, gaseous), which depending on its concentration may be toxic to these
191 organisms [12].

192 One of the first studies to determine the effects of nitrogen on natural environments and its association with other
193 parameters was performed by Trussel (1972) [30], concluding that in these environments the concentration of ammonia

194 very rarely reaches lethal levels. For these levels to be reached, high pH (>9.0), temperature (>260°C) and low
195 oxidation potential values must occur simultaneously.

196 3.6 Phosphorus

197 Phosphorus is an important element in fundamental processes of living metabolism, such as: energy storage (it
198 forms an essential fraction of the ATP molecule) and cell membrane structure (through phospholipids) [12].

199 In most continental waters phosphorus is the main limiting factor of its productivity. Moreover, it has been pointed
200 as the main responsible for the artificial eutrophication of these ecosystems [28; 31; 32].

201 All forms of phosphorus present in natural waters, both ionic and complexed, are in phosphate form which can be
202 divided into five main forms: particulate phosphate (P - particulate), dissolved organic phosphate (P - dissolved
203 organic), dissolved inorganic phosphate or orthophosphate or reactive phosphate (P-ortho), total dissolved phosphate (P
204 - total dissolved) and total phosphate (P - total) [3].

205 From a limnological point of view, all forms (also called fractions) and phosphate are important, however, P-ortho
206 assumes greater relevance because it is the main form of phosphate assimilated by aquatic plants. Thus, its
207 quantification in limnological evaluations becomes indispensable [33].

208 Phosphate present in continental aquatic ecosystems originates from natural and artificial sources. Among natural
209 sources, drainage basin rocks, especially apatite, constitute the basic source of phosphate for continental aquatic
210 ecosystems. The release of phosphate from the crystalline form of the primary rock minerals occurs through its
211 breakdown by weathering [12].

212 Other natural factors that allow phosphate input can be pointed out, such as: particulate matter present in the
213 atmosphere and phosphate resulting from the decomposition of organisms of allochthonous origin. The most important
214 artificial sources of phosphate are domestic and industrial sewage and industrial particulate matter contained in the
215 atmosphere [12].

216 3.7 Biochemical oxygen demand

217 The reduction in dissolved oxygen concentration is the main problem of water pollution and occurs because of the
218 respiratory activity of bacteria to stabilize organic matter. Therefore, the presence of organic matter in water can be
219 assessed by measuring oxygen consumption. To indicate the oxygen consumption or demand required to stabilize
220 organic matter in water, the parameters BOD (Biochemical Oxygen Demand) and COD (Chemical Oxygen Demand)
221 are used [22].

222 The difference between BOD and COD is in the type of stabilized organic matter: while BOD refers exclusively to
223 mineralized organic matter by activity of microorganisms, COD also encompasses the stabilization of organic matter by
224 chemical processes based on the fact that some organic compounds are oxidized by oxidizing chemicals considered
225 strong in acidic environments. Therefore, the COD value is always higher than the BOD value.

226 As COD values are generally higher than those found for BOD, in laboratory tests, COD is determined first, and
227 the results serve as guidelines for establishing dilutions for BOD testing. The relationship between COD and BOD
228 values indicates the amount of organic matter that can be biologically stabilized. This demand is referred conventionally
229 to a period of five days, since the complete stabilization of organic matter requires a longer time, and at a temperature of
230 20°C [34].

231 Both BOD and COD are expressed in mg/L. Most values found in the national literature report organic loading in
232 terms of BOD, which is, between the two, the most commonly used parameter to characterize domestic sewage and
233 industrial effluents [35].

234 In unpolluted natural environments, BOD concentration is low (1 mg/L to 10 mg/L) and can reach much higher
235 values in bodies of water subject to organic pollution, usually due to the receipt of domestic sewage or animal breeding.
236 Effluents from industries that process organic matter (dairy, breweries, refrigerators) have BOD values in the order of
237 tens or even hundreds of grams per liter [22].

238 3.8 Temperature

239 The change in water temperature can be caused by natural (mainly solar) or anthropogenic (industrial waste and
240 machine cooling water) sources and can significantly influence the aquatic environment and its biota by interfering with
241 the speed of chemical reactions, biochemical and physiological activities of organisms and the solubility of substances
242 [36].

243 The thermal regime of a body of water is a crucial factor for ecosystem quality. In Brazil, aquatic environments
244 generally have temperatures in the range of 20°C to 30°C. However, in colder regions, such as in the south of the
245 country, winter water temperatures can reach values between 5°C and 15°C, reaching, in some cases, the freezing point
246 [22].

247 Temperature also influences other physicochemical parameters of water, such as surface tension and viscosity.
248 Aquatic organisms are affected by temperatures outside their thermal tolerance limits, which impacts their growth and
249 reproduction. The higher the temperature, the lower the surface tension that serves as a substrate for the life of small
250 organisms [37; 38]. Viscosity reflects its flow resistance inversely proportional to temperature, which means that a hot
251 water is less viscous than a cold water, interfering, among other factors, in the mobility of small organisms and
252 sedimentation of suspended particles [39].

253 Due to the high specific heat of water, aquatic environments are quite stable with respect to temperature. Despite
254 the thermal stability, the water bodies present temperature variations throughout the day (to a greater or lesser extent
255 depending on the local morphoclimatic domain) and the seasons of the year. However, discharging effluent at high
256 temperatures can have a significant impact on water bodies. Regarding water for human consumption, high
257 temperatures increase the prospects of rejection to use [40].

258 3.9 Turbidity

259 Turbidity is the parameter that represents the occurrence of light passing through water due to the presence of
260 suspended particles, giving it a turbid appearance [41] expressed by means of turbidity units (also called Jackson or
261 nephelometric units) [22]. The aesthetically undesirable aspect caused by increased turbidity is derived from the
262 presence of suspended particles in different dimensions, ranging from colloids to coarse suspensions, including organic
263 detritus, algae, bacteria and plankton that compromise light passage, causing their scattering and absorption.

264 Aquatic biological communities are strongly affected by turbidity variation, since their increase, for example,
265 reduces photosynthesis from submerged rooted vegetation and algae, and may therefore suppress fish productivity [12].
266 In addition, high turbidity directly affects domestic, industrial and recreational water use.

267 In Brazilian water bodies, due to geological characteristics and high rainfall rates in most drainage basins, turbidity
268 is particularly high because rainfall on erodible soils can carry clay particles, silt, sand, rock fragments and metal oxides
269 from the soil.

270 In lakes and dams, where there is little runoff, turbidity is lower. The natural turbidity of water in these
271 environments ranges from 3 to 500 Units (NTU). According to CONAMA Resolution (2005), for class 3 freshwater
272 bodies, the turbidity limit is 100 NTU (turbidity unit). In filtered water, turbidity assumes a sanitary and not merely
273 aesthetic indicator function. For potability purposes, turbidity should be less than one unit, ensuring the effectiveness of
274 filtration, disinfection and absence of pathogens.

275 3.10 Total (solid) residue

276 Residues or Solids are all materials suspended or dissolved in water, in domestic or industrial waste. Total residue
277 is the matter that remains after evaporation, drying or calcination of the water sample for a given time and temperature.
278 In general, the drying, calcining and filtration operations define the various fractions of solids present in water (total,
279 suspended, dissolved, fixed and volatile solids) [42].

280 Total solids can be classified into total suspended solids and total dissolved solids, and according to the thermal
281 treatment performed on the sample, the solids can be further fragmented into “fixed” and “volatile” terms, which are
282 determined gravimetrically, an analytical procedure based on the difference between dry mass and wet mass, in relation
283 to the sample volume disposed in the test at different temperatures [43; 44].

284 Although turbidity and total solid parameters are intuitively associated, they are not at all equivalent. A stone, for
285 example, placed in a glass of clean water, gives to that environment a high concentration of total solids, but its turbidity
286 may be practically null [22].

287 4. Conclusion

288 Lakes and their basins are unique and integrated units, so all processes and problems are interconnected. Thus, this
289 ecosystem must be managed without jurisdictional limitation and various policy instruments need to be introduced.

290 It is understood that this management should be institutionalized through policies and systems, which define and
291 give effect to the roles of governments, civil society and the private sector, at the national, regional and global levels.

292 Water in nature presents a series of dissolved compounds that adhere to it through the atmosphere, soils, rocks and
293 the decomposition of organic matter. These compounds are responsible for the characteristics of water even in its
294 natural condition [3]. Because of the long response time, impacts are sustained for a long time, solutions take time, and
295 changes are often gradual and invisible. Long-term monitoring commitment based on reliable geochemical analysis is
296 required.

297 The complex response dynamics in these ecosystems also make responses to impacts unpredictable, making
298 scientific involvement in linking problems to their cause's indispensable, with a long-term and politically strong
299 commitment. Based on this, water classification systems fall into two main types: those concerned with the amount of
300 pollution present and those concerned with communities living on macroscopic or microscopic organisms.

301 The assignment of a numerical value based on physicochemical variables should consider the conformity of local
302 water and climate characteristics and their adaptations to the analysis, although presenting satisfactory results, masks a
303 variety of conditions that occur in a water course, e.g. presence of pathogenic microorganisms. Although its synthesis
304 capacity makes it an important instrument with strong public utility, it is noteworthy that the WQI represents a quality
305 index, not an instrument for evaluating public service.

306 One of the current major challenges in environmental quality management is to reach a balance between the
307 activities that generate economic development and the preservation of the environment. This challenge has been
308 translated through the concept of sustainable development, which aims to meet the needs of the current generation,
309 without compromising the ability to meet the needs of future generations [3].

310 Integrated watershed management should seek to address challenges for sustainable use of lake and river resources.
311 The term 'integrated management' refers to the simultaneous consideration of all relevant management factors,
312 including scientific and socioeconomic ones. The scientific elements are highlighted in their importance as input of
313 theoretical support and technical information, the socioeconomic are major components of lake management and are,
314 therefore, factors that should be aligned with the sustainable use of services of that ecosystem and a given basin [10].

315 Although there is much to be celebrated regarding the legal framework and the progress made, huge challenges
316 remain to be met in order to achieve effective integrated management. These challenges refer to issues of great
317 complexity that will still deserve enormous efforts to overcome integration of water management with environmental
318 management; integration of quantity and quality aspects; integration of surface and groundwater management; and
319 especially the integration of water resources policy with sectoral policies [3]. Efforts should be made, however, towards
320 an innovative approach generating new management practices and policies.

321 The systemic model of water resources management adopted for Brazil is still in the process of improvement,
322 despite the great and current advances achieved since the beginning of its implementation. The challenges pointed out,
323 at no time, question the model's postulates and premises, nor do they advocate their revision. They only point out that,
324 given the country's diversity and the different levels of water resources commitment along its territory, it is reasonable
325 to think of multiple views and adaptation care that are not yet given [3].

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333 References

- 334 [1] Silva, J. A. A.; Nobre, A. D.; Manzatto, C. V.; Joly, C. A.; Rodrigues, R. R.; Skorupa, L. A.; Nobre, C. A.; Ahrens,
335 S.; May, P. H.; Sá, T. D. A.; Cunha, M. C.; Rech Filho, E. L. (2011) O Código Florestal e a ciência: Contribuições
336 para o diálogo. *São Paulo: Sociedade Brasileira Para o Progresso da Ciência, SBPC; e Academia Brasileira de*
337 *Ciências, ABC*, 124p.
- 338 [2] Gorgulho, S. Recursos Hídricos: gestão de lagos. (2003) *Folha do Meio Ambiente*.
339 <http://www.folhadomeio.com.br/publix/fma/folha/2003/02/garrido.html>.
- 340 [3] Porto, R. L. L. (2012) *Fundamentos para gestão da água*. São Paulo. s.n. 232 p.
- 341 [4] Lobato da Costa, F. J. (2003) Estratégias de gerenciamento dos recursos hídricos no Brasil: áreas de cooperação com
342 o Banco Mundial. *Série Água Brasil n. 1*. Banco Mundial. Brasília.
- 343 [5] Brasil. Ministério do Meio Ambiente; Agência Nacional de Águas. (2012) *Pacto nacional pela gestão das águas:*
344 *construindo uma visão nacional*. Brasília.
- 345 [6] BRASIL. Ministério do Meio Ambiente; Agência Nacional de Águas. Programa das Nações Unidas para o Meio
346 Ambiente. (2007) *Gestão dos recursos hídricos no Brasil: evolução e “estado da arte”*. GEO BRASIL: recursos
347 hídricos. Brasília.
- 348 [7] Rohde, S. M. (2004) *Geoquímica ambiental e estudos de impacto*. 2ed. São Paulo. Sigmus Editora, 157 p.
- 349 [8] Martins, R.; Porto, M. (2014) *Decaimento e mistura de poluentes no meio ambiente*.
350 [Http://www.fctb.br/phd/phd2460/Aulas/PHD%202460%20Aula%2011.pdf](http://www.fctb.br/phd/phd2460/Aulas/PHD%202460%20Aula%2011.pdf).
- 351 [9] Aleksander-Kwaterczak, U.; Zdechlik, R. (2016) Hydrogeochemical characteristics of interstitial water and
352 overlying water in the lacustrine environment. *Environmental Earth Sciences*, v. 75, n. 20, p. 1352.
- 353 [10] Cudishevitch, C., Fontanetto, R., Oliveira, S. (2012) Gestão e políticas públicas para bacias hidrográficas.
354 *Academia Brasileira de Ciência*. http://www.abc.org.br/article.php?id_article=2168
- 355 [11] MOE, S. J.; HAANDE, S.; COUTURE, R. M. Climate change, cyanobacteria blooms and ecological status of lakes:
356 A Bayesian network approach. *Ecological Modelling*, v. 337, 2016, p. 330-347.
- 357 [12] Esteves, F.A. (2011) *Fundamentos de limnologia*. 3. ed. Rio de Janeiro: Interciência, 826 p.
- 358 [13] Shadrin, N. V. (2017) Hypersaline lakes as polyextreme habitats for life. *Introduction to Salt Lake Sciences*.
359 Science Press, Beijing, China, p. 173-178.
- 360 [14] Ballester, M.; Santos, J. E. (2001) Biogenic gases (CH₄, CO₂ and O₂) distribution in a wetland system. *Oecologia*
361 *Australis*, v. 9, n. 1, p. 21-32.
- 362 [15] Duncan, S. et al. (2019) Dynamics of residential water-soluble organic gases: Insights into sources and
363 sinks. *Environmental science & technology*, v. 53, n. 4, p. 1812-1821.
- 364 [16] Hanson, P. C., Carpenter, S. R., Armstrong, D. E., Stanley, E. H., & Kratz, T. K. (2006) Lake dissolved inorganic
365 carbon and dissolved oxygen: changing drivers from days to decades. *Ecological Monographs*, v. 76, n. 3, p.
366 343-363.

- 367 [17] Breitburg, D. et al. (2018) Declining oxygen in the global ocean and coastal waters. *Science*, v. 359, n. 6371, 2018,
368 p. eaam7240.
- 369 [18] Hachich, E. M. et al. (2012) Comparison of thermotolerant coliforms and *Escherichia coli* densities in freshwater
370 bodies. *Braz. J. Microbiol.* São Paulo, v. 43, n. 2, p. 675-681.
371 <http://dx.doi.org/10.1590/S1517-83822012000200032>.
- 372 [19] Cerqueira, D. A.; Horta, M. C. S. (1999) Coliformes fecais não existem. In: *Congresso Brasileiro de Engenharia*
373 *Sanitária e Ambiental*, p. 1239-1244.
- 374 [20] Hodge, J. et al. (2016) Assessing the Association between Thermotolerant Coliforms in Drinking Water and
375 Diarrhea: An Analysis of Individual-Level Data from Multiple Studies. *Environmental health perspectives*, v. 124,
376 n. 10, p. 1560-1567.
- 377 [21] SMITH, M., MINTEER, C., LAKE, E.C., WHEELER, G.S., TIPPING, P.W. (2016) Indirect ecological effects in
378 invaded landscapes: Spillover and spillback from biological control agents to native analogues. *Ecological Society*
379 *of America (ESA)*, https://eco.confex.com/eco/2016/preliminaryprogram/abstract_59794.htm
- 380 [22] Brasil. (2006) Ministério da Saúde. Secretaria de Vigilância em Saúde. *Vigilância e controle da qualidade da água*
381 *para consumo humano*. Brasília.
- 382 [23] Mirzaei, N. et al. (2015) Survey of effective parameters (water sources, seasonal variation and residual chlorine) on
383 presence of thermotolerant coliforms bacteria in different drinking water resources. *International Journal of*
384 *Pharmacy and Technology*, v. 7, n. 3, p. 9680-9689.
- 385 [24] Robertson, W. et al. (1998) Evaluation of a rapid method for *E coli* and thermotolerant coliforms in recreational
386 waters. *Water science and technology*, v. 38, n. 12, p. 87.
- 387 [25] Maurer, C. P. et al. (2015) Adenovirus, enterovirus and thermotolerant coliforms in recreational waters from Lake
388 Guaíba beaches, Porto Alegre, Brazil. *Journal of water and health*, p. wh2015277.
- 389 [26] Tennant, C. J. (2016) *Analysis of Inorganic Carbon and pH in the Western Arm of Lake Superior*. [Tese de
390 Doutorado]. University of Minnesota.
- 391 [27] Thienemann, A. (1918) Lebensgemeinschaft und Lebensraum. *Naturwissenschaftliche Wochenschrift* v. 17, n. 20,
392 p. 282-290.
- 393 [28] Biudes, J. F. V.; Camargo, A. F. (2008) M. Estudos dos fatores limitantes à produção primária por macrófitas
394 aquáticas no Brasil. *Oecologia Brasiliensis*, v. 12, n. 1, p. 2.
- 395 [29] Ochocka, A.; Pasztaleniec, A. (2016) Sensitivity of plankton indices to lake trophic conditions. *Environmental*
396 *Monitoring and Assessment*, v. 188, n. 11, p. 622.
- 397 [30] Trussell, R. P. (1972) The Percent Un-Ionized Ammonia in Aqueous Ammonia Solutions at Different p H Levels
398 and Temperatures. *Journal of the Fisheries Board of Canada*, v. 29, n. 10, p. 1505-1507.
- 399 [31] Schindler, D. W. et al. (2016) *Reducing phosphorus to curb lake eutrophication is a success*.
- 400 [32] Dodds, W. K.; Smith, V. H. (2016) Nitrogen, phosphorus, and eutrophication in streams. *Inland Waters*, v. 6, n. 2,
401 p. 155-164.

- 402 [33] Gong, L. J., Yang, X. F., Xiong, B. X., Li, G. P., & Chen, X. L. (2012) Study on Nitrogen, Phosphor and Chemical
403 Oxygen Demand of Different Categories of Aquaculture Lakes by Means of Principal Component Analysis, Factor
404 Analysis and Cluster Analysis. In: Advanced Materials Research. *Trans Tech Publications*, p. 369-377.
- 405 [34] American Public Health Association (APHA). (2012) Standard methods for the examination of water and
406 wastewater. 22 nd ed., Washington, *American Public Health Association Pub.*, 1935 p.
- 407 [35] Brandão, C. J. et al. (2011) Guia nacional de coleta e preservação de amostras: água, sedimento, comunidades
408 aquáticas e efluentes líquidos. *São Paulo: CETESB*.
- 409 [36] Verones, F. et al. (2010) Characterization factors for thermal pollution in freshwater aquatic
410 environments. *Environmental science & technology*, v. 44, n. 24, p. 9364-9369.
- 411 [37] Allen, D. C.; Vaughn, C. C. (2011) Density - dependent biodiversity effects on physical habitat modification by
412 freshwater bivalves. *Ecology*, v. 92, n. 5, p. 1013-1019.
- 413 [38] Min, K. R.; Zimmer, M. N.; Rickard, A. H. (2010) Physicochemical parameters influencing coaggregation between
414 the freshwater bacteria *Sphingomonas natatoria* 2.1 and *Micrococcus luteus* 2.13. *Biofouling*, v. 26, n. 8, p.
415 931-940.
- 416 [39] Hanafiah, M. M. (2013) *Quantifying effects of physical, chemical and biological stressors in life cycle assessment*.
417 [Tese de Doutorado].
- 418 [40] Zlatanović, L.; Van Der Hoek, J. P.; Vreeburg, J. H. G. (2017) An experimental study on the influence of water
419 stagnation and temperature change on water quality in a full-scale domestic drinking water system. *Water research*,
420 v. 123, p. 761-772.
- 421 [41] Gloria, L. P.; Horn, B. C.; Hilgemann, M. (2017) Avaliação da qualidade da água de bacias hidrográficas através
422 da ferramenta do Índice de Qualidade da Água-IQA. *Revista Caderno Pedagógico*, v. 14, n. 1.
- 423 [42] Kindlein, C. P. (2010) *Determinação do teor de nitratos e nitritos na água de abastecimento do Município de Nova*
424 *Santa Rita*. v. 200. http://biblioteca.unilasalle.edu.br/docs_online
- 425 [43] Lamparelli, M. C. (2014) *Graus de trofia em corpos d'água do estado de São Paulo: avaliação dos métodos de*
426 *monitoramento*. [Tese de Doutorado]. Universidade de São Paulo.
- 427 [44] Garcez, L. N. (2004) Manual de procedimentos e técnicas laboratoriais voltado para análises de águas e esgotos
428 sanitário e industrial. *São Paulo: Escola Politécnica de Universidade de São Paulo*, 105p.